Hydrological simulation as subside for management of surface water resources at the Mortes River Basin

Simulação hidrológica como subsídio para a gestão dos recursos hídricos superficiais na Bacia do Rio das Mortes

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ABSTRACT

Hydrological models are tools which describe processes and allow predicting the results of making management decisions, and are an important water resource management tool, especially for small-sized watersheds. This work had as an objective to test different calibration strategies and apply the SWAT model for hydrological simulation of the Mortes River Basin, MG. We evaluated 6 fluviometric stations, with drainage area between 272 and 6070 km² for purposes of water resource management. The evaluation of the model performance was conducted by using Nash-Sutcliffe coefficients (C_{NS}) and percent bias (P_{BMS}). The results obtained from the statistic indices applied in the analysis of the model performance qualified the SWAT hydrological model as a dequate for streamflow simulation in the Mortes River Basin. The adopted strategies attest to the applicability of the model as a management tool for water resources planning for other small-sized watersheds without data, in order to plan for rational water use.

Index terms: Environmental management; SWAT; drainage area; calibration.

RESUMO

Os modelos hidrológicos são ferramentas que descrevem processos e permitem antecipar as ocorrências e tomar decisões, dessa forma, apresentam-se como uma importante ferramenta de gestão dos recursos hídricos, principalmente no contexto de bacias hidrográficas de pequeno porte. Assim, objetivou-se com este trabalho abordar diferentes estratégias de calibração e teste do modelo SWAT na simulação hidrológica da bacia hidrográfica do Rio das Mortes, MG. Foram avaliados 6 postos fluviométricos, com área de drenagem entre 272 e 6.070 km², para fins de gestão dos recursos hídricos. A avaliação do desempenho do modelo foi realizada usando os coeficientes de Nash-Sutcliffe (C_{NS}) e percentual de viés (P_{BIAS}). Os resultados obtidos dos índices estatísticos aplicados na análise de desempenho do modelo, qualificam o modelo hidrológico SWAT como adequado na simulação da vazão para a bacia hidrográfica do Rio das Mortes. As estratégias adotadas atestaram a aplicabilidade do modelo como ferramenta de gestão e planejamento dos recursos hídricos para bacias menores e sem dados, possibilitando o uso racional da água na bacia.

Termos para indexação: Gestão ambiental; SWAT; área de drenagem; calibração.

INTRODUCTION

The economic and technological development and accelerated population growth have generated a series of conflicts over natural resources, with water use being the most common of these. The recent problems of water supply and the increasing pressure of occupation and degradation of river basins have generated research demand regarding the changes in the hydrological dynamics of watersheds.

Understanding of water dynamics in a watershed is of fundamental importance to water resource planning and management programs. According to its behavior, it can be assessed if the watershed is either in balance or in the degradation process (Pinto et al., 2013).

Knowledge of the interrelationships associated with natural phenomena involving the hydrologic cycle has enabled the development of mathematical models that describe the water dynamics in natural ecosystems. Recently, the detailing of the physical characteristics by the linkage of geoprocessing and remote sensing techniques has led to improved modeling of hydrologic systems. In this context, hydrological models of conceptual and distributed bases present themselves as an important tool for water resource management, with the possibility of applying them for streamflow forecasting, assessment of climate scenarios, soil use and studies on water availability (Viola et al., 2012), among others.

Among the numerous models that have been used for watershed planning and management, the SWAT model, (acronym for "Soil and Water Assessment Tool") stands out (Arnold et al., 1998). The model was developed by the Agricultural Research Service (USDA-ARS) and by Texas A&M University. It is a public domain tool designed to predict the impact of soil management practices, sediment and agrochemicals upon watersheds (Neitsch et al., 2011).

SWAT is one of the hydrological models most widely used in the world (Piniewski; Okruszko, 2011), being accepted internationally as a robust and versatile model of watersheds (Bressiani et al, 2015a). In Brazil, some studies were conducted by applying the SWAT as tool for hydrological studies, highlighting: Paim and Menezes (2009), in the state of Santa Catarina; Andrade, Mello and Beskow (2013), Durães, Mello and Naghettini (2011), Lelis and Calijuri (2010), Lelis et al. (2012), Melo Neto et al. (2014), Pinto et al. (2013), and Rodrigues et al. (2015), all in the state of Minas Gerais; Lessa et al. (2013), Galharte, Villela and Crestana (2014), in the state of São Paulo; Aragão et al. (2013), in the state of Sergipe; Fukunaga et al. (2015), and Pereira et al. (2014) in the state of Espírito Santo; Bonumá et al. (2015) in the state of Rio Grande do Sul; and Bressiani et al. (2015b) in the state of Ceará. These studies have shown satisfactory results for SWAT use as a tool in hydrologic simulation, aiming at the multiple daily demands of water resource management.

The ability of a hydrological model to simulate the hydrological cycle in a watershed is evaluated through both calibration and validation processes (White; Chaubey, 2005). These processes are critical since the models present a number of uncertainties involved in a phenomenon as complex as the hydrologic cycle.

The calibration of a hydrological model for a watershed has been commonly developed by applying data from a single fluviometric station. This is inevitable in a region where there is only one station, or even in small-sized watersheds. However, as the hydrologic variables (such as streamflow) have elevated temporal and spatial variabilities, the development of the modeling with data from only one station may not be able to adequately characterize the phenomenon in time and space in a watershed (Noor et al., 2014). Thus, some studies have been conducted comparing the calibration of a single station with that of a set of stations (Piniewski; Okruszko, 2011; Noor et al, 2014).

Due to the shortage of information of the hydrometric network for water resource management studies in the state of Minas Gerais, the need for studies of the behavior of streamflow is warranted. These studies may allow improving existing fluviometric information, by extrapolating results to sites which possess data either insufficient or inexistent, as is often the case of watersheds with small-sized drainage areas located in upstream regions. Viola et al. (2009) pointed out that hydrological models can be useful for better management of water resources, especially in the context of small-sized watersheds. According to Soares et al. (2010), the determination of the water availability in small watersheds, generally, devoid of streamflow measurements, constitutes a major problem in the application of the water resource management instruments.

In this case, there is a demand for studies on calibration strategies of hydrological models that can provide information for planning and management of water resources, especially for smaller basins and without data.

The Mortes River Basin is characterized by various mining projects, metallurgical activities and industrialization. The Mortes River represents one of the most important tributaries of the Grande River; its source is in the Mantiqueira range, near the city of Barbacena and the river mouth ends at the Funil Dam - Funil Hydroelectric Power Station (FHS). The relief presents itself is rolling and with significant extents already compromised by agricultural use, by the mining activities and the presence of gullies and rills (Curi; Chagas; Giarola, 1994) which present great sediment production potential.

This study was conducted to address different calibration strategies for the SWAT model in the hydrologic simulation of the Mortes River Basin, MG, by evaluating 6 fluviometric stations with drainage area between 272 and 6070 km² for water resource management purposes.

MATERIAL AND METHODS

Study area

The Mortes River Basin (MRB) is located in the southeastern region of the State of Minas Gerais, Brazil (Figure 1). The MRB presents a 6070 km² drainage area, it lies upstream to the Funil Hydroelectric Station (FHS) and is one of the contributors to this station and to the very Furnas hydroelectric station downstream at the dam.

The watershed under study possesses the climate types Cwb and Cwa, according to the Köppen classification, characterized by mild and humid summers and cold and dry winters, with annual average temperature of 18 °C and annual average rainfall of around 1500 mm.



Figure 1: Geographical localization of the Mortes River Basin.

The SWAT model

The model "Soil and Water Assessment Tool" (SWAT) (Arnold et al., 1998) has been applied in many studies worldwide, especially in research related to water balance, soil management, transport of sediments, nutrients and pesticides, water quality, climate change and land use. SWAT is a semi-distributed model, continuous in time, and capable of simulation over long periods of time (Neitsch et al., 2011; Gassman et al., 2007). In simulation, the watershed is divided into sub-basins, which are then subdivided into hydrologic response units (HRU's). Each HRU's represent a homogeneous combination of characteristics of land use, soil type and slope classes (Neitsch et al., 2011). In this study, the version of SWAT2012 was utilized, through the ArcGIS 10.2 software interface, named ArcSWAT2012.

For hydrologic modeling of the watershed, SWAT requires a set of information such as: a digital terrain model (DEM), map of land use and soil cover and a soil class map. In addition, it is also necessary to feed the model with meteorological data such as rainfall, solar radiation, wind speed, relative humidity and maximum and minimum temperatures (Figure 2).

For application of the model, a hydrometeorological database with daily-step was developed. It was acquired from the Hydrological Information System of the National Water Agency (ANA), rainfall and fluviometric historical series (HIDROWEB, 2014), in addition to the records of the INMET station meteorological elements (BDMEP, 2014) presented in (Figure 2a and Table 1).

Vector topographic maps of the Systematic Mapping at the scale of 1:50,000 provided by the Brazilian Institute of Geography and Statistics (IBGE) were used for the DEM. Such maps were interpolated to generate the DEM allowing generating a regular grid with a spatial resolution of 30 vs. 30 m, this was the resolution used in the study. To obtain the DEM, these interpolation procedures were performed from tools available in ArcGIS 10.2 software interface. The altitude in the watershed ranges from 842 to 1,387 m (Figure 2b).

The soil map was obtained from the soil survey available in FEAM (2010), in the scale of 1: 650,000. In the watershed, the classes of Latosol soil occur in 45.1% of the area, Cambisol in 38.2% of the area; Argisol in 15.4% of the area; rocky outcrop in 0.7% of the area and Neosol in 0.6% of the area (Figure 2c).

For the soil use and cover mapping, images from the OLI sensor (Operational Land Imager) aboard the Landsat 8 satellite from 2013, with a spatial resolution of 30 meters and temporal passing every 16 days (Figure 2d) were utilized. The adopted classification method was object-oriented classification which aimed to split the image into homogeneous regions through its spectral and spatial attributes, followed by classification of these objects split into established land cover classes.

For modeling, the runoff curve number method (SCS-CN) for representation of the direct surface runoff process and the Penman-Monteith method for evapotranspiration were selected (Neitsch et al., 2011).

Sensitivity, calibration, validation analyses and uncertainty analysis

The sensitivity, calibration and validation analyses and uncertainty analysis were carried out in SWAT-CUP version 5.1.6 software (Abbaspour, 2015) using the algorithm "Sequential Uncertainty Fitting Procedure" Ver. 2 (SUFI-2) (Abbaspour; Johnson; van Genuchten, 2004; Abbaspour et al., 2007).

The sensitivity analysis of the SWAT model is conducted by means of the combination of the method of the Latin Hypercube sampling (LH) and of the One-Factor-At-a-Time structure (OAT) (Griensven et al., 2006). The sensitivity analysis determines which are the parameters most important to a particular watershed, which directs the set of parameters to be utilized in the subsequent calibration process. For this purpose, 16 parameters (Table 2) considered by the literature to be the most sensitive for streamflow simulation were used (Abbaspour et al., 2007; Andrade; Mello; Beskow, 2013; Durães; Mello; Naghettini, 2011; Lelis et al., 2012; Melo Neto et al., 2014). The calibration procedure used all six fluviometric stations: Vila do Rio das Mortes (272 km²), Campolide (569 km²), Usina (SJDR) (643 km²), Porto do Elvas (828 km²), Barroso (1040 km²) and Ibituruna (6070 km²).

With the purpose of evaluating the SWAT performance as to the number of stations applied in calibration and respective drainage area (Da), different strategies for calibration and test of the model were carried out:



Figure 2: Hydroclimatologic monitoring stations (a), digital elevation model (b), soils map (c) and land use and soil cover map (d) of the MRB for the application of the SWAT model.

| Code | Name | Type of station | Observation period | |
|----------|----------------------------------|-----------------|--------------------|--|
| 83689 | Barbacena | Weather | 1993-2005 | |
| 2044050 | São Tiago | Pluviometric | 1993-2005 | |
| 2143005 | Campolide | Pluviometric | 1993-2005 | |
| 2143006 | Barroso | Pluviometric | 1993-2005 | |
| 2143008 | Ibertioga | Pluviometric | 1993-2005 | |
| 2143009 | Usina Barbacena | Pluviometric | 1993-2005 | |
| 2144000 | Bom Sucesso | Pluviometric | 1993-2005 | |
| 2144002 | Porto Tiradentes | Pluviometric | 1993-2005 | |
| 2144009 | Porto do Elvas | Pluviometric | 1993-2005 | |
| 2144020 | Usina de São João Del Rei (SJDR) | Pluviometric | 1993-2005 | |
| 2144023 | Ibituruna | Pluviometric | 1993-2005 | |
| 2144024 | Vila do Rio das Mortes | Pluviometric | 1993-2005 | |
| 61135000 | Ibituruna | Fluviometric | 1993-2005 | |
| 61115000 | Usina SJDR | Fluviometric | 1993-2000 | |
| 61105000 | Porto do Elvas | Fluviometric | 1993-2000 | |
| 61122000 | Vila do Rio das Mortes | Fluviometric | 1993-2000 | |
| 61090000 | Barroso | Fluviometric | 1993-2000 | |
| 61085000 | Campolide | Fluviometric | 1993-2000 | |

Table 1: Basic information of the utilized stations.

1) Calibration (1993 to 2000) and validation (2001 to 2005) utilizing exclusively the fluviometric data from the Ibituruna station, which was near the outlet of the studied watershed. Under this situation, the other five stations were utilized to test the capacity of the model to preview the runoff in upstream sub-basins, in the period of 1993 to 2000. The years of 1993 and 1994 were used as periods for warming up of the model since the early conditions for the model are unknown such as conditions of early moisture of the soil in the watershed. Warming up period is utilized for, when beginning the simulation, so that the variables would be free of the influences of early starting conditions (Zhang et al., 2007).

In addition our strategy, we compared the observed and simulated daily data to average streamflow, and the values of minimum and maximum streamflows for the calibration and validation period for the MRB.

2) Calibration considering all the available fluviometric stations.

3) Calibration considering each fluviometric station individually. It was aimed in this situation to compare the

quality of the individual fit of each station, with the joint calibration expressed in item (2).

4) It was evaluated the quality of the calibration resulting the combination of different fluviometric stations according to their drainage area, in the period of 1993 to 2000. They were:

4.1) Calibration using two stations with largest area (Da = 6070 and 1040 km^2): under this situation the other four stations were utilized to test the predictive capacity of the model in representing areas within smaller drainage areas;

4.2) Calibration based upon the station of largest drainage area ($Da = 6070 \text{ km}^2$) and smallest area ($Da = 272 \text{ km}^2$) and in this situation, a performance test of the model was conducted in the four monitored sections not employed in the calibration process;

5) Calibration considering only the data of the fluviometric station of smallest drainage area ($Da = 272 \text{ km}^2$). In this approach, the other five stations were utilized to test the predictive capacity of the model in representing areas with a larger drainage area.

| Parameter | Description | Process | Initial value | Calibrated value |
|-----------|--|------------------------------|------------------|---------------------|
| ESCO | Soil water evaporation compensation coefficient | Evapotranspiration | 0.80 | 0.739 |
| CN2 | Initial curve-number for moisture conditions II | Surface runoff | 55 to 92 | -0.06%* |
| ALPHA_BF | Base flow recession coefficient (days) | Groundwater | 0.048 | 0.004 |
| GW_DELAY | Time interval for aquifer recharge (days) | Groundwater | 31 | 68.15 |
| GWQMN | Water limit in the shallow aquifer for occurrence of base flow (mmH20) | Groundwater | 1000 | 1279 |
| CANMX | Maximum amount of water intercepted by vegetation (mmH20) | Evapotranspiration | 0 | 13.23 |
| CH_K2 | Effective hydraulic conductivity in the main channel (mm.h ⁻¹) | Transmission water losses | 0 | 1.166 |
| CH_N2 | Manning's "n" value for the main channel | Surface runoff | 0.014 | 0.299 |
| EPCO | Water absorption coefficient by plants | Groundwater | 1 | 0.103 |
| GW_REVAP | Water rise coefficient to the saturated zone | Groundwater | 0.02 | 0.048 |
| REVAPMN | Soil water limit for occurrence of capillary rise to the saturated zone (mmH20) | Groundwater | 750 | 1368 |
| SOL_AWC | Available water capacity of the soil (mmH2O. mmsoil ⁻¹) | Soil water | 0.142 to 0.175 | +0.005%* |
| SOL_K | Soil saturated hydraulic conductivity | Soil water | 5.18 to 67.32 | +0.024%* |
| SURLAG | Surface runoff delay coefficient (days) | Surface runoff | 2 | 2.26 |
| CH_N1 | Manning's "n" value for the secondary channel | Surface runoff | 0.014 | 0.264 |
| CH_K1 | Effective hydraulic conductivity in the tributary channel (mm.h ⁻¹) | Transmission water losses | 0 | 3.06 |

Table 2: Influent parameters in the streamflow simulating process by the SWAT and their initial values and final calibrated value.

*parameters altered based on their initial values varying percentages depending on the class, land use and soil layer referred.

The tests were conducted as follows: in each hydrologic simulation was considered a fluviometric station or the combination of stations in the calibration step. Once calibrated, the parameter values were applied to hydrologic simulation in other fluviometric stations. For each strategy calibration five iterations with 600 simulations each were performed. After each iteration, the sensitivity of the 16 parameters was determined, having as a criterion parameters presenting the value p < 0.05 were selected and had their ranges reduced for the next iteration. Such a procedure finds support in the fact that such parameters present greater sensitivity since values close to 0 indicate a high significance (Abbaspour, 2015).

For practical application of the strategies adopted, permanence curves were developed for the sub-sections of the sub-basins, having as a basis the best strategy tested. To assess the quality of the model adjustment to the observed data, the Nash-Sutcliffe coefficient (CNS - Equation 1) (Nash; Sutcliffe, 1970) and the percent bias of the simulated streamflow in relation to that observed (PBIAS - Equation 2) were utilized (Gupta; Sorooshian; Yapo, 1999).

$$C_{NS} = 1 - \frac{\sum_{i=1}^{n} (Q_{OBSi} - Q_{SIMi})^{2}}{\sum_{i=1}^{n} (Q_{OBSi} - \overline{Q}_{OBS})^{2}}$$
(1)

$$P_{BLAS} = \left(\frac{\sum_{i=1}^{n} (Q_{OBSi} - Q_{SIMi})}{\sum_{i=1}^{n} (Q_{OBSi})}\right).100$$
 (2)

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Where, Q_{OBSi} and Q_{SIMi} corresponds to the observed and simulated streamflow, respectively, on day *i* (m³ s⁻¹), and \overline{Q}_{OBS} corresponds to the average observed streamflow, in (m³ s⁻¹), and *n* corresponds to the number of events.

According to Zhang et al. (2007), the value of the Nash-Sutcliffe coefficient indicates the adjustment of simulated data to those observed on the straight 1: 1 line and can range from - ∞ to 1. To evaluate the C_{NS} performance, the criterion proposed by Moriasi et al. (2007) was adopted, which presents the following rating for this coefficient using a daily simulation step: C_{NS} > 0.65, the model is considered very good; 0.54 < C_{NS} < 0.65, the model is considered good and between 0.50 and 0.54, satisfactory.

The PBIAS measures the average trend of the simulated data to be either larger or smaller than their equivalents observed (Moriasi et al., 2007). The classification by Van Liew et al. (2007) was adopted to evaluate the accuracy of the model in application of SWAT using daily intervals. According to this criterion: $|P_{BIAS}| < 10\%$, very good; $10\% < |P_{BIAS}| < 15\%$, good; $15\% < |P_{BIAS}| < 25\%$, satisfactory and $|P_{BIAS}| > 25\%$, the model is inadequate.

In the SUFI-2 algorithm, the uncertainties of the input parameters are represented as uniform distributions, while the output uncertainties of the model are expressed as probability distributions of 95%. These are calculated at the 2.5% and 97.5% levels of the cumulative distribution of output variables obtained by the propagation of the uncertainties of the parameters using Latin Hypercube sampling (LH). This is referred to as the 95% prediction uncertainty, or 95PPU. These 95PPUs are the model outputs in a stochastic calibration approach. It is important to realize that there is not a single signal representing model output, but rather an envelope with possible solutions expressed by 95PPU, generated by certain parameter ranges. The SUFI-2 starts by assuming a large parameter uncertainty, so that the measured data initially falls within the 95PPU. In sequence, decrease this uncertainty in steps until two rules are satisfied: (1) the 95% prediction uncertainty (95PPU) supports more than 90% of the measured data, and (2) the average distance between the upper part (at 97.5% level) and the lower part (at 2.5% level) is lesser than the standard deviation of the measured data (Abbaspour; Johnson; van Genuchten, 2004; Abbaspour et al., 2007).

To quantify the fit of the simulation, expressed at 95PPU and observation expressed as a single signal (with any errors associated with it), two statistics were used: P-factor and R-factor (Abbaspour; Johnson; van Genuchten, 2004; Abbaspour et al., 2007). The P-factor is the percentage of measured data involved by the result of the modeling, in other words, it is the 95PPU and varies from 0 to 1. The value 1 indicates 100% bracketing of the measured data within model prediction uncertainty (or being, a perfect model simulation considering the uncertainty). The R-factor, on the other hand, is the ratio of the average width of the 95PPU band by the standard deviation of the measured data, that is, the average thickness of the 95PPU band. For variable streamflow, values of P-factor > 0.7 and of R-factor of around 1 are recommended, considering thus the calibration as being adequate (Abbaspour, 2015, Abbaspour et al., 2007).

RESULTS AND DISCUSSION

Strategy 1: Behavior of the model at the Mortes River Basin

A tool used in the evaluation of the accuracy of hydrological models consists in analyzing visually the simulated hydrogram in relation to that observed. Figure 3 presents both the simulated and observed hydrograms and the daily hietograms for the Mortes River Basin at Ibituruna (6070 km²) during calibration and validation periods, respectively. In general, the simulated streamflow adjusted well to the observed ones, i.e., the model managed to capture the frequent oscillations inherent to long-term continuous simulation.

In performing a visual analysis of the hydrograms, one realizes, in some situations, difficulty of simulating maximum streamflow. According to Pereira et al. (2014), streamflow peaks are naturally difficult to model, which is related to spatial and temporal representation problems of the most intense rains. Viola et al. (2009) add that the simulation step is another factor limiting to the simulation of the maximum streamflow events, since the sub-basins present reduced concentration time in relation to the main watershed. This difficulty was also observed by other authors, with the SWAT model, such as Andrade, Mello and Beskow (2013), Durães, Mello and Naghettini (2011), Melo Neto et al. (2014), Pereira et al. (2014), and Pinto et al. (2013).

Another point that can be observed in the hydrographs is a relative difficulty for the model to represent the recession after events with peak streamflow. According to Melo Neto et al. (2014) this is because there is a delay of the model to represent the recession curve decay, simulating a greater volume in these periods compared to the observed series.



Figure 3: Daily hydrograms observed and simulated and respective hietogram, results of the precision statistics resulting from SWAT calibration and validation, and 95% prediction uncertainty (95PPU).

In Figure 3 are also presented the results of the precision statistics applied to evaluate the performance of the SWAT model in the hydrological simulation of the Mortes River Basin calibrated and validated exclusively with Ibituruna station streamflow data. By analyzing the values, we observed good results in all the evaluated parameters. For the results of the uncertainty analysis, good results were obtained for both factors, i.e. values of P-factor > 0.7 and values of R-factor close to 1. That set of values allows classifying calibration on the basis of that strategy as adequate (Abbaspour et al., 2007).

The values of the C_{NS} coefficient for the calibration step was of 0.76, which is classified as very good and for the validation step, it was of 0.62, which is classified as good, according to the classification by Moriasi et al. (2007) (Figure 3). It is usual for the statistical indices of the validation period to be inferior to the indices for the calibration period, since the values of the parameters are specifically optimized for the calibration period (Fukunaga et al., 2015; Lelis et al., 2012; Pinto et al., 2013). The validation period can present different conditions which cause the calibrated parameters to be different from the ideal (Fukunaga et al., 2015).

According to Durães, Mello and Naghettini (2011) the validation period of the simulation process is critical, allowing the definitive characterization of the applicability of the model. That happens due to the fact that in this phase the model is applied to a new database, considering the parameters obtained during the calibration period. Bonumá et al. (2015) points out that good validation results support the model's utility in predicting other situations such as alternative scenarios of soil use and management, future climates and generation of simulated historical series.

Thus, since the C_{NS} coefficient is the precision statistics most applied and important to hydrologic simulation, it can be concluded that the model will produce good results for this watershed, proving an important tool for planning and management of surface water resources.

Analyzing the performance of the model based on the results of P_{BIAS} (Figure 3), the values found indicated an average underestimation of 6.0% of the streamflow in the calibration process and overestimation of 5.5% in the validation phase and these values being considered as very good according to the classification of Van Liew et al. (2007).

Durães, Mello and Naghettini (2011) evaluated the performance of the SWAT model in the hydrologic simulation of the Paraopeba River watershed, MG, with 10,200 km² drainage area and obtained de C_{NS} values equal to 0.79, both in calibration and validation, the model being defined as appropriate and good. Melo Neto et al. (2014) evaluated the performance of the SWAT model in estimating the streamflow in two watersheds with different spatial scales (32 km² and 4.7 km²) in the Upper Grande River region, south of Minas Gerais. The authors found C_{NS} values between 0.58 and 0.71 in the calibration phase and of 0.46 in the validation phase for mesoscale. By contrast, in the microscale, the model obtained inadequate performance with values of the statistical coefficients below the limits recommended in the literature.

Pereira et al. (2014) evaluated the performance of the SWAT model in the hydrologic simulation of a watershed located in Domingos Martins town, ES, with 943 km² drainage area. The authors found $C_{\rm NS}$ values of

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0.65 and 0.70 for the calibration and validation periods, respectively, which provides the basis for the model to be classified as satisfactory. Fukunaga et al., 2015 evaluated the SWAT capacity to model daily streamflow in the Itapemirim River watershed, ES, with a 2237 km² drainage area. The authors found C_{NS} values of 0.75 and 0.67 for the calibration and validation periods, respectively, and concluded that SWAT hydrologic model provided satisfactory results to the streamflow simulation of in the studied watershed.

In order to test the performance of the SWAT model calibrated with data of the Ibituruna fluviometric station, the simulation of the behavior of the upstream streamflow, data of the Usina (SJDR), Porto do Elvas, Vila do Rio das Mortes, Barroso and Campolide stations were used.

In Table 3 are presented the results of accuracy statistics obtained by the calibration based on strategy 1, i.e., simulation for upstream sub-basins with smaller drainage areas. The values obtained for the $C_{\rm NS}$ coefficient ranged from 0.28 (Porto Elvas station) to 0.50 (Barroso station), the latter being the only station with value of the $C_{\rm NS}$ coefficient classified as satisfactory.

Table 3: Values of precision statistics obtained in the SWAT model performance test in the stations Usina (SJDR), Porto do Elvas, Vila do Rio das Mortes, Barroso, Campolide, using the parameters calibrated for the lbituruna station.

| Da* | Stations | C _{NS} | P _{BIAS} |
|------|------------------------|-----------------|-------------------|
| 272 | Vila do Rio das Mortes | 0.38 | 15.4 |
| 569 | Campolide | 0.45 | 42.7 |
| 643 | Usina (SJDR) | 0.35 | -13.1 |
| 828 | Porto do Elvas | 0.28 | -1.5 |
| 1040 | Barroso | 0.50 | -1.3 |

*Drainage area in km²; $\rm C_{\rm NS}$ - Nash-Sutcliffe coefficient; $\rm P_{\rm BIAS}$ - percent bias

As the $C_{\rm NS}$ coefficient is an important statistical index in the evaluation of hydrologic modeling, even though most of the stations have not presented $C_{\rm NS}$ values considered satisfactory, these values can provide a starting point to the future studies addressing upstream tests of the studied test section.

The differences observed between calculated and observed streamflow, i.e., the percent bias (P_{BIAS}) ranged from 42.7% (Campolide station) to -1.3% (Barroso station), indicating high underestimation of the model for the Campolide station and slight overestimation of the model at the Barroso and Porto dos Elvas stations (Table 3). According to the classification proposed by Van Liew et al. (2007), the model showed very good performance P_{BIAS} (<10%) at the Barroso, Porto do Elvas stations; good (10 < P_{BIAS} <15%) at Usina (SJDR); satisfactory 15% < $|P_{BIAS}| < 25\%$ in Vila do Rio das Mortes and non-satisfactory ($P_{BIAS} \ge 25\%$) at Campolide. These results, together with a satisfactory performance obtained with the statistical coefficient, allowed concluding that the model had an adequate performance upstream using the calibration data at the Barroso station.

Table 4 lists the comparison between observed and simulated daily data to average streamflow, and values of minimum and maximum streamflows for the calibration and validation period for the MRB. We can observe smaller errors for average streamflow, denoting an adequate quality of the hydrological model in the simulation for the MRB. Greater errors were found for the maximum and minimum streamflows, which are expected, keeping in mind that the main aim of this study was the streamflows continuous simulation. Results of the same quality were found by Pereira et al. (2014) and Viola et al. (2009).

Considering the results, it can be inferred that the parameters calibrated at the watershed outlet may be used in future studies to represent the hydrological behavior of sub-basins, in the extent of the watershed under study. This information is of importance to the management of water resources, especially in large watersheds and with a low density of fluviometric stations.

Table 4: Comparison between observed (Obs.) and simulated (Sim.) daily data to average streamflow, and minimum and maximum values of streamflows for the MRB.

| Period | Q _{ave} * | | | Q _{max} | | Q _{min} | | | |
|-------------|--------------------|-------|-------|------------------|-------|------------------|------|------|-------|
| | Obs. | Sim. | Error | Obs. | Sim. | Error | Obs. | Sim. | Error |
| Calibration | 123.6 | 116.1 | 6.1 | 1299 | 1888 | 45.3 | 40.2 | 39.2 | 2.5 |
| Validation | 120.3 | 126.9 | 5.5 | 692.4 | 959.1 | 38.5 | 35 | 52.9 | 51.1 |

*Q_{ave} = average streamflow (m³ s⁻¹); Q_{max} = maximum streamflow (m³ s⁻¹); Q_{min} = minimum streamflow (m³ s⁻¹); Error (%).

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Strategies 2 and 3: Calibration considering all the stations jointly and individually

In Figure 4 are shown the results of the statistical accuracy obtained in the calibration of all the stations jointly, as well as of the individual calibration. Analyzing Figures 4a and b, it is observed, in general, for the results of the uncertainty analysis, a growing behavior of values P-factor and R-factor as a function of the drainage area of the stations, with the P-factor > 0.7. For the R-factor, the best fits were also obtained for the stations with a larger drainage area, that is, when the value of R-factor is close to 1, the calibration is regarded as appropriate (Abbaspour et al., 2007). The different results found reflect the uncertainties involved in modeling natural systems.

According to Abbaspour (2015) one of the greatest sources of error in the input data are the rainfall information, being especially relevant in areas with occurrence of convective or orographic precipitations, with high spatial and percent variability. According to Bouraoui et al. (2004), an inadequate representation of the required rainfall data or lack of rainfall collection

stations in the watershed makes it impossible to collect data that are spatially representative for this variable. According to Piniewski and Okruszko (2011), for use of the SWAT model on smaller spatial scales, which is an important demand of the water resource management, it is necessary to provide better input data and, in particular, use a monitoring network of the rainfall regime sufficiently detailed to describe the variability of rainfall events.

Figure 4c shows the graphical representation of the values of the $C_{\rm NS}$ coefficient of the calibration of all stations jointly with each individual calibrated station. Analyzing this figure, it can be seen that for both types of calibration, $C_{\rm NS}$ values showed an increasing behavior in function of the drainage area of the evaluated stations. For calibration of the jointly stations, the $C_{\rm NS}$ values ranged from 0.39 to 0.73. For the Barroso station the $C_{\rm NS}$ values was 0.59 which is classified as good, and for the Ibituruna station, 0.73, which is classified as very good, according to criteria used in this study (Moriasi et al., 2007).

In the calibration based on individual stations, the Barroso station obtained a $C_{\rm NS}$ of 0.70, the Ibituruna station, 0.76, which are classified as very good. For the



Figure 4: Graphical representation of the values of the precision statistics, P-factor (a), R-factor (b), C_{NS} - Nash-Sutcliffe coefficient (c), and P_{BIAS} - percent bias (d) obtained in the calibration considering all the fluviometric stations jointly and individually.

Campolide station we obtained a value of 0.57, which is classified as good (Figure 4c). Despite the $C_{\rm NS}$ value for the Campolide station having been satisfactory, when the other results of precision statistics are evaluated (PBIAS, P-factor and R-factor), it is found that they are inadequate. Thus, this demonstrates the importance of evaluating the overall results, characterized by several statistical coefficients and uncertainty analysis, and not by an isolated precision statistical value. According to Abbaspour (2015) there is a close relationship between calibration and uncertainty, thus in studies involving stochastic calibration, the evaluation of this set of information is important.

Analyzing the percent bias (P_{BIAS}) for calibration of the stations jointly, it is observed that most of the stations were within the class values considered adequate, except for the Campolide station since it showed a value of 41.3%. a result that is considered inappropriate. For the Porto dos Elvas, Barroso and Ibituruna stations calibrations were classified as very good, that is, values of $|P_{\text{BIAS}}| < 10\%$. For the Usina (SJDR) and Vila do Rio das Mortes stations, calibration is classified as good, that is, values in the range of $10\% < |P_{BIAS}| < 15\%$ (Figure 4d). The stations (SJDR) and Mortes Rio Village calibration were classified as good, that is, values in the 10% range $\leq |P_{BIAS}| \leq 15\%$ (Figure 4d). For the station calibration individually, the same behavior for most stations was observed, except for the Campolide station which had a value of 29.2%, which was inadequate. For all other stations evaluated, calibration was classified as very good, that is, values of $|P_{BIAS}| < 10\%$ (Figure 4d).

Piniewski and Okruszko (2011) studied the behavior of the drainage area with regard to calibration and validation of 12 sub-basins ranging from 355 to 1,657 km², in a watershed located in Northeastern Poland. For both evaluations, the authors concluded that the model was better in large watersheds and that the model should not be used for areas lesser than 600 km², under the conditions studied by the authors.

Thus, for the watershed under study, it can be concluded that the calibration of fluviometric stations in an individual way, when compared with the calibration of the stations jointly resulted into a small gain in the quality of $C_{\rm NS}$ coefficient adjustment for all drainage areas. In summary, for the watershed under study, the SWAT model performed better in the calibration of the station in an individual mode than for calibration of the stations jointly.

Noor et al. (2014) evaluated the calibration performance of a station compared with the calibration of several stations inside the same watershed in Iran with drainage areas ranging from 805 to 48 km². The authors concluded that the calibration based on a single station and based on several stations generated acceptable de $C_{_{NS}}$ values.

According to Daggupati et al. (2015) use of a single location to calibrate large watersheds can result in calibrated parameters that represent the mean of the characteristics in the watershed as a whole or present a combination of over or underestimated values that result into poor spatial accuracy within the watersheds. This may be undesirable for simulations of larger watersheds that are more spatially heterogeneous.

Therefore, when working with large watersheds, it becomes important to use more than one fluviometric station in the calibration process to assess whether the use of one or more stations results in differences in the performance of the simulation.

Strategy 4: SWAT performance associated with the number of stations applied in the calibration and respective drainage areas

Aiming to evaluate the performance of SWAT as to the number of fluviometric stations applied in the calibration and respective watersheds we conducted combinations of fluviometric sections for calibration and model testing. The sections of Ibituruna, Barroso and Vila Rio das Mortes were considered for the calibration. While the other fluviometric stations in the watershed were used to test the model's performance.

The results are shown in Figure 5 where it can be seen that the model is sensitive to the settings analyzed. When comparing the results obtained from the jointly calibration of the two stations with larger drainage areas (Ibituruna Da = 6070 km² and Barroso 1040 km²) with the combination of the stations including larger and smaller drainage area (Ibituruna Da = 6070 km² and Vila do Rio das Mortes Da = 272), it was observed by the C_{NS} coefficient values that there was a gain in the fit quality for the other stations. This was especially noticiable for the Porto do Elvas (828 km²) and Barroso (1040 km²), which presented C_{NS} values of 0.45 and 0.69, respectively. That shows that the model presents a good predictive ability for smaller drainage areas, when combining the larger drainage area stations in the calibration process.

Still analyzing Figure 5, it is observed by the results obtained that the SWAT model presents difficulty of simulation smaller drainage areas. According to Piniewski and Okruszko (2011) some of the possible reasons that can explain this performance of the model are that in small upstream watersheds, the local rainfall response is very fast, while in downstream watersheds, the largest streamflow results in smoother hydrographs, the smaller the catchment area, more errors in the input data such as rainfall, land use, in other words, the spatial heterogeneity becomes smaller and that super- and under-estimation of streamflow in small upstream watersheds may to do averaging in the downstream watershed. Uzeika et al. (2012) also report that the main limitation of SWAT to represent the variables in daily time scale is related to short and fast rain events, which occur in small watersheds.

Considering the importance of managing water resources inside a watershed, we can highlight the results obtained when combining stations with larger drainage areas. These analyses are important, especially considering the Brazilian conditions, where there is a lack of information of the hydrometric network many times, with insufficient or no data, as in the case of watersheds with small drainage areas, located in upstream regions. Fukunaga et al. (2015) pointed out that SWAT hydrological model is a useful tool to be applied to simulate streamflow for tropical watersheds under low availability of environmental data, especially meteorological data.

Strategies 5: Calibration considering the station of smallest drainage area

Aiming to evaluate the predictive capacity of SWAT in representing areas with larger drainage area, we proceeded using the calibration of the Vila do Rio das Mortes station, which presents the smallest drainage area of the watershed ($Da = 272 \text{ km}^2$). The other fluviometric stations present in the watershed were utilized to test the performance of the model.

In analyzing the results presented in Figure 6 it can be observed that calibration with the station with the smallest drainage area resulted in good performance of the model, except fir values of the $C_{\rm NS}$ coefficient obtained, for the Campolide, Barroso and Ibituruna stations, which presented values of $C_{\rm NS}$ coefficient of 0.50, 0.61 and 0.65 respectively, being classified as "satisfactory" for the first station and as "good" for other two. These results show that the model presents good predictive capacity to represent areas with larger drainage areas, therefore, a potential tool for management, even when the calibration watershed is smaller, as in this case.

Development of the permanence curves

The curve of permanence provides the frequency with which determined streamflow is exceeded, having as the main functionality a determination of minimum streamflow of reference for a location, for water resource management (Viola et al., 2012).

In order to demonstrate the applicability of the work for water management purposes, permanence curves of streamflow for the sub-basins studied were developed, using strategy 4.1 as a basis which was the



••••• Ibituruna and Barroso

Figure 5: Graphical representation of the values of Nash Coefficient (C_{NS}) obtained in the evaluation of the calibration and test of the SWAT model in function of the drainage area of the stations.

one which presented the best performance of the model, when associated with the two stations with the larger drainage area of the watershed. A reference value often extracted from this curve is the Q90%, which represents the streamflow surpassed or equaled by 90% of the time. For the curves presented in Figure 7, the simulated values of Q90% were of 1.68, 2.02, 4.04, 7.47 and 6.90 m³ s⁻¹ for the Vila do Rio das Mortes, Campolide, Usina (SJDR), Porto do Elvas and Barroso sub-basins, respectively.

Andrade, Mello and Beskow (2013) showed that when it comes to the estimate of reference streamflow for a location there is a series of limitations for small watersheds, since the availability of hydrological data of streamflow is limited or practically inexistent at this scale in Brazil.

For Fukunaga et al. (2015) the SWAT model has the potential to be used as a useful tool for water resource management, especially when applied to two of the most important instruments of Brazilian water resource policy which are the concession of water resource use rights and charging for the use of water resources.

In that way, the results obtained in this work demonstrate the importance of the development of methodologies which seek to give support for water resource management, highlighting its need in small watersheds without data.



Figure 6: Graphical representation of the Nash Coefficient values (C_{NS}) obtained in the evaluation of the calibration and test SWAT model in function of the drainage area of the stations, considering as the calibration section the stations of the smallest area in the watershed.



Figure 7: Permanence curves simulated by SWAT for the Mortes River sub-basins.

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CONCLUSIONS

The results obtained from the statistical indices applied in the model performance analysis found the hydrological SWAT model as appropriate for the simulation of the streamflow in the Mortes River watershed. The SWAT model was sensitive to the size of the drainage area, showing good results when combining the larger drainage area stations in the calibration process. The adopted strategies confirmed the applicability of the model as a management and planning tool of water resources for smaller watersheds without data. The use of the SWAT model in the Mortes River watershed can provide a better understanding of hydrological processes and can provide support for planning and decision making related to water resource management, especially for smaller watersheds without data, enabling the rational use of water in the watershed.

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